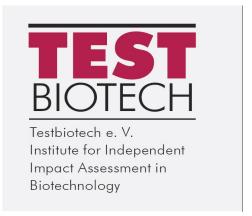
TESTBIOTECH Background 08 - 12 -2019

Testbiotech comment on EFSA's assessment of genetically engineered soybean MON87751 x MON87701 x MON87708 x MON89788 for food and feed uses, under Regulation (EC) No 1829/2003 (application EFSA-GMO-NL-2016-128) by Bayer/Monsanto



Christoph Then & Andreas Bauer-Panskus & Juliana Miyazaki

Introduction

The GMO panel assessed the four-stacked soybean event MON87751 x MON87701 x MON87708 x MON89788 derived from crossing genetically engineered soybean events. The parental soybeans have undergone previous assessment (EFSA, 2019a). The soybean contains genes conferring resistance to two herbicides:

- MON87751 expressing the insecticidal proteins Cry1A.105 and Cry2Ab2, with Cry1A.105 being synthetic without natural template;
- MON87701, expressing the insecticidal protein Cry1Ac;
- MON89788 expressing CP4 EPSPS protein for tolerance to glyphosate-containing herbicides;
- MON87708 expressing dicamba mono-oxygenase (DMO), for tolerance to the herbicide dicamba.

Consequently, the stacked GE soybean is resistant to two groups of complementary herbicides (glyphosate and dicamba) and produces three insecticidal proteins. The herbicides can be applied in combination or individually. Implementing Regulation 503/2003 was applied in this case.

1. Molecular characterisation

The process of genetic engineering involved several deletions and insertions in the parental soybean plants. In order to assess the sequences encoding the newly expressed proteins or any other open reading frames (ORFs) present within the insert and spanning the junction sites, it was assumed that the proteins that might emerge from these DNA sequences would raise no safety issues; therefore, no detailed investigations were carried out in this regard. Furthermore, other gene products, such as dsRNA from additional open reading frames, were not assessed. Thus, uncertainties remain about other biologically active substances arising from the method of genetic engineering and the newly introduced gene constructs.

Furthermore, it seems EFSA (2019b) did not request specific data on the place of the insertion of the additional constructs in the genome. Instead, EFSA proposes that such data is not needed ("No requirements are laid down in the Implementing Regulation as to provide the exact location of the events in the plant genome.") However, data are necessary to consider possible position effects. Therefore, EFSA should have requested much more detailed investigation into potential biologically active gene products, position effects and changes in metabolic pathways.

In regard to expression of the additionally inserted genes, Implementing Regulation 503/2013 requests "Protein expression data, including the raw data, obtained from field trials and related to the conditions in which the crop is grown" (in regard to the newly expressed proteins)."

However, there are three reasons why the data presented do not represent the conditions in which the plants will be grown: (1.1) the field trials were not conducted in all relevant regions where the soybeans will be cultivated, and no extreme weather conditions were taken into account; (1.2) the field trials did not take into account current agricultural management practices; (1.3.) only one transgenic variety was included in the field trials.

1.1

Environmental stress can cause unexpected patterns of expression in the newly introduced DNA (see, for example, Trtikova et al., 2015). More specifically, Fang et al. (2018) showed that stress responses can lead to unexpected changes in plant metabolism inheriting additional EPSPS enzymes. However, the expression of the additional enzymes was only measured under field conditions in the US.

As mentioned by the experts of Member States, higher application rates of the complementary herbicides can cause stress reactions in the plants and impact gene expression (EFSA, 2019b). Therefore, the plants should also have been tested in large soybean producing countries in South America, which not only differ in soil and climate but also in agricultural practice. For example, there are publications showing higher rates of glyphosate applications in South America compared to the US (Benbrook, 2016).

Whatever the case, the plants should have been subjected to a much broader range of defined environmental conditions and stressors (which, for example, have to be expected under ongoing climate change) to gather reliable data on gene expression and functional genetic stability.

1.2

Due to increased weed pressure, it has to be expected that these plants will be exposed to high and also repeated dosages of glyphosate alone and / or in combination with dicamba. Higher applications of herbicides will not only lead to a higher burden of residues in the harvest, but may also influence the expression of the transgenes or other genome activities in the plants. As mentioned by the experts of Member States, higher application rates of the complementary herbicides can cause stress reactions in the plants and impact gene expression (EFSA, 2019b). However, this aspect was ignored in the EFSA risk assessment. While currently, 'on top' glyphosate applications at an average rate of 3 to 4 kg / ha and an average overall rate of 6 to 7 kg /ha (USDA, 2019) can be expected in the US and even more in South America (see, for example, Bombardi, 2016), the amount of glyphosate used in the field trials was just 0,87 kg a.e./ha, which is close to the lowest limit of application recommended by the company (EFSA 2019b).

A statement made by EFSA (2019b) indicates that the design of the field trials should avoid major differences in the application of the herbicides:

"The complementary herbicides are kept at a similar application rate across sites: indeed, for the experimental treatments to be comparable between different locations, the application rate should not differ too strongly between them."

This statement is in direct contradiction to the requirements of Implementing Regulation 503/2013. If there are any problems in regard to comparability, EFSA should request more data which are necessary to establish a dose-response curve; using specific amounts of pesticide applied during the field trials. These data would allow the comparison and interpretation of the relevant findings.

EFSA should have requested the applicant to submit data from field trials with the highest dosage of the complementary herbicides that can be tolerated by the plants, including repeated spraying and the application of each of the relevant herbicides alone and in combination. The material derived from those plants should have been assessed by using omics techniques to investigate changes in the gene activity of the transgene, as well as in the natural genome of the plants.

1.3.

It is known that the genomic background of the variety can influence the expression of the inserted genes (see, for example, Trtikova et al., 2015). In this case, it was shown that the expression of the (naturally occurring) allergen Gly m 4 is lower in the stacked event compared to its conventional comparator. Further significant differences concern (naturally occurring) phyto-estrogens (daidzein and genistein) and the concentration of the newly expressed Cry1A.105, DMO and EPSPS enzymes. This indicates influences from the process of stacking and the resulting overall genomic background of the stacked event. Therefore, EFSA should have requested data from the parental plants to be grown in parallel. Further additional data from several varieties, including those cultivated in South America, would have been necessary.

The material derived from the plants should have been assessed by using 'Omics-techniques' to investigate changes in the gene activity of the transgene and the plants genome, as well as changes in metabolic pathways and the emergence of unintended biological active gene products. Such indepth investigations should not depend on findings indicating potential adverse effects, they should always be necessary to come to sufficiently robust conclusions to inform the next steps in risk assessment.

2. Comparative analysis (for compositional analysis and agronomic traits and GM phenotype) Implementing Regulation 503/2013 requests:

"In the case of herbicide tolerant genetically modified plants and in order to assess whether the expected agricultural practices influence the expression of the studied endpoints, three test materials shall be compared: the genetically modified plant exposed to the intended herbicide; the conventional counterpart treated with conventional herbicide management regimes; and the genetically modified plant treated with the same conventional herbicide management regimes."

"The different sites selected for the field trials shall reflect the different meteorological and agronomic conditions under which the crop is to be grown; the choice shall be explicitly justified. The choice of non-genetically modified reference varieties shall be appropriate for the chosen sites and shall be justified explicitly."

However, the data that were presented do not represent anticipated agricultural practices, or the different meteorological and agronomic conditions where the crop is to be grown. The following three reasons can be given: (2.1) the field trials were not conducted in all relevant regions where the soybeans will be cultivated, and no extreme weather conditions were taken into account; (2.2) the field trials did not take current agricultural management practices into account; (2.3) only one transgenic variety was included in the field trials.

2.1

Field trials for the compositional and agronomic assessment of the stacked soybeans were only conducted in the US, but not in other relevant soybean production areas such Brazil, Argentina, Paraguay or Uruguay. As stated in the EFSA opinion (2019a), "No exceptional weather conditions were reported at any of the selected field trial sites."

It is not acceptable that EFSA failed to require further studies, e.g.

- Just one field trial was conducted that lasted more than one season. Thus, based on current data, it is hardly possible to assess site-specific effects.
- Further, no data were generated representing more extreme environmental conditions, such as those caused by climate change.

More specifically, Fang et al. (2018) showed that stress responses can lead to unexpected changes in plant metabolism due the production of the additional EPSPS enzymes. However, no experiments were requested to show to which extent specific environmental conditions will influence plant composition or agronomic characteristics.

As mentioned by the experts of Member States, higher application rates of the complementary herbicides can cause stress reactions in the plants and impact gene expression (EFSA, 2019b). Therefore, the plants should have also been tested in large soybean producing countries in South America, which not only differ in soil and climate but also in agricultural practice. For example, there are publications showing higher rates of glyphosate applications in South America compared to the US (Benbrook, 2016).

Whatever the case, the plants should have been subjected to a much broader range of defined environmental conditions and stressors (which, for example, have to be expected under ongoing climate change) to gather reliable data on plant composition and phenotypical characteristics.

2.2

Due to high weed pressure in many soybean growing regions, it has to be expected that these plants will be exposed to higher amounts and repeated dosages of the herbicides. It has to be taken into account that the herbicides can be sprayed in combination or individually at high dosages and repeatedly. These agricultural practices have to be taken into account to assess whether the expected agricultural practices will influence the expression of the studied endpoints. Higher applications of herbicides will not only lead to a higher burden of residues in the harvest, but may also influence the expression of the transgenes or other genome activities in the plants. As mentioned by the experts of Member States, higher application rates of the complementary herbicides can cause stress reactions in the plants and impact gene expression (EFSA, 2019b). However, this requirement was mostly ignored by EFSA and the company: the herbicides were only sprayed in combination, each just once, at an early stage of vegetation and at comparably low dosages: The amount of glyphosate used in the field trials was just 0,87 kg a.e./ha, which is close to the lowest limit of application recommended by the company (EFSA, 2019b).

Available publications show that the complementary herbicides are sprayed at much higher dosages and repeatedly onto the GE soybeans: on its product label Monsanto recommends spraying with about 7 kg (a.i.)/ha (Monsanto, 2017), with up to three applications during cultivation. Official figures from the USDA data base show that up to 6-7 kg (a.i.)/ha of glyphosate can be expected in soybean cultivation, including pre- and post-emergence applications (USDA, 2019). Data from South America show that even higher amounts are possible (Avila-Vazquez et al., 2018).

From the data that is available, it has to be assumed that the specific patterns of complementary herbicide applications will not only lead to a higher burden of residues in the harvest, but may also influence the composition of the plants and agronomic characteristics. This aspect was ignored in the EFSA risk assessment.

It is known that soybeans contain many biologically active substances, e.g. estrogens, allergens and anti-nutritional compounds, which may interact with trait-related characteristics and act as stressors. Changes in the composition of these components may not only be triggered by the process of genetic engineering, but also by interactions with the complementary herbicides. For example, Zobiole et al. (2012) and also Bøhn et al. (2014) found that glyphosate application can cause significant changes in soybean plant constituents. More specifically, Zobiole et al. (2012) applied glyphosate at three different dosages (800, 1200 and 2400 g/ha), which resulted in dose-correlated changes in plant agronomic performance and plant composition.

A statement made by EFSA (2019b) indicates that the design of the field trials should avoid major differences in the application of the herbicides:

"The complementary herbicides are kept at a similar application rate across sites: indeed, for the experimental treatments to be comparable between different locations, the application rate should not differ too strongly between them."

This statement is in direct contradiction to the requirements of Implementing Regulation 503/2013. If there are any problems in regard to comparability, EFSA should request more data necessary to establish a dose-response curve; using specific amounts of pesticide applied during the field trials. These data would allow the comparison and interpretation of the relevant findings.

It also should be taken into account that a mixture of all the complementary herbicides will not always be used in the fields where the soybeans are cultivated; in some cases, just one of them will be used. This might lead to an increase in dosages of the respective complementary herbicides. The choice of herbicide will depend on the price of the herbicide formulations, the respective weed problem and regional agricultural practices. For example, it can be expected that in Argentina, Brazil and the US, there will be different prices, different herbicide formulations and varying regimes of herbicide applications under which the soybean is cultivated. None of these specific agronomic practices were considered in the design of the field trials or in EFSA risk assessment.

EFSA should have requested the company to submit data from field trials with the highest dosage of the complementary herbicides that can be tolerated by the plants, including repeated spraying with each active ingredient individually as well as in combination.

2.3

It is known that the genomic background of the variety can influence the expression of the inserted genes (see, for example, Trtikova et al., 2015). In this case, it was shown that the expression of natural allergen Gly m 4 is lower in the stacked event compared to its conventional comparator. Further significant differences concern isoflavones or so called phyto-estrogens (daidzein and genistein), and the concentration of the newly expressed Cry1A.105, DMO and EPSPS enzymes. This indicates influences from the process of stacking and resulting overall genomic background of the stacked event. Therefore, EFSA should have requested data from the parental plants to be grown in parallel as well as additional data from several varieties, including those cultivated in South America.

The material derived from the plants should have been assessed by using 'Omics-techniques' to investigate changes in the gene activity of the transgene and the plants genome, as well as changes in metabolic pathways and the emergence of unintended biological active gene products. Such indepth investigations should not depend on findings indicating potential adverse effects, they should always be necessary to come to sufficiently robust conclusions to inform the next steps in risk assessment.

2.4

Only data from a low number of agronomic parameters (8), were subjected to statistical analysis in accordance with EFSA guidance; 2 of these were found to be significantly different in the stacked plants compared to their conventional counterparts. Against the backdrop of significant differences even in this small data set, EFSA should have requested much more data (see also above).

Compositional analysis of 53 endpoints in the grains revealed many (and partly major) statistically significant differences: in comparison to their conventional counterparts 25 endpoints were significantly different in plants not sprayed with the complementary herbicides and 16 in the stacked plants that were sprayed. One of them ("Gly m 4") protein indicated major differences between the transgenic stack and its comparator.

As shown above, the data show a much lower number of significant findings in the plant composition and the phenotypical characteristics if the plants were sprayed with the complementary herbicides. This indicates that metabolic pathways might have been impacted by the application of the complementary herbicide. This should have been investigated in more detail.

Therefore, EFSA should have requested further tests (toxicological data, repeated spraying with higher herbicide dosages or exposure to a wider range of environmental conditions). Furthermore, the plant material should have been assessed in more detail by using omics techniques to investigate changes in plant composition and agronomic characteristics.

But instead of assessing in more detail the overall pattern of changes in plant components, their causes and possible impacts, EFSA only assessed the observed changes in isolation. This approach turns the comparative approach into a trivial concept of assessing bits and pieces, and ignores questions concerning the overall safety of the whole food and feed.

More in-depth investigations should not depend on findings indicating adverse effects, they should always be necessary to come to sufficiently robust conclusions to inform the next steps in risk assessment. Even if changes taken as isolated data might not directly raise safety concerns, the overall number of effects and their clear significance has to be taken as a starting point for much more detailed investigations.

Based on the available data, no final conclusions can be drawn on the safety of the plants. In any case, the data do not fulfill the requirements of Implementing Regulation 503/2013.

Toxicology

Implementing Regulation 503/2013 requests:

- "Toxicological assessment shall be performed in order to:
- (a) demonstrate that the intended effect(s) of the genetic modification has no adverse effects on human and animal health;
- (b) demonstrate that unintended effect(s) of the genetic modification(s) identified or assumed to have occurred based on the preceding comparative molecular, compositional or phenotypic analyses, have no adverse effects on human and animal health;"
- "In accordance with the requirements of Articles 4 and 16 of Regulation (EC) No 1829/2003, the applicant shall ensure that the final risk characterisation clearly demonstrates that:
- (a) the genetically modified food and feed has no adverse effects on human and animal health:"

There were many significant changes in plant composition and agronomic characteristics, Furthermore, several uncertainties were identified in the feeding studies with the parental plants. Nevertheless, testing of the whole stacked plant (feeding study) was not requested. Even if changes taken as isolated data might not directly raise safety concerns, the overall number of effects should have been considered as a starting point for much more detailed investigation of their potential health impacts.

Beyond that, the residues from spraying were considered to be outside the remit of the GMO panel. However, without detailed assessment of these residues, no conclusion can be drawn on the safety of the imported products: due to specific agricultural practices in the cultivation of these herbicideresistant plants, there are, e.g. specific patterns of applications, exposure, occurrence of specific metabolites and emergence of combinatorial effects that require special attention (see also Kleter et al., 2011).

More detailed assessment is also in accordance with pesticide regulation that requires specific risk assessment of imported plants if pesticide usage in the exporting countries is different compared to EU usage. In this regard, it should be taken into account that EFSA (2019c) explicitly stated that no conclusion can be drawn on the safety of residues from spraying with glyphosate occurring in genetically engineered plants resistant to this herbicide.

The analysis of the toxicity data for glyphosate and dicamba indicate a higher toxicity if the two herbicides are combined (Reuter, 2015). EFSA should have at least requested data on the combined toxicity of the residues from spraying with the complementary herbicides.

Further, there is a common understanding that commercially traded formulations of glyphosate, such as Roundup, can be more toxic than glyphosate itself. Therefore, the EU has already taken measures to remove problematic additives known as POE tallowamine from the market. Problematic additives are still allowed in those countries where the genetically engineered plants are cultivated. The EU Commission has confirmed the respective gaps in risk assessment:

"A significant amount of food and feed is imported into the EU from third countries. This includes food and feed produced from glyphosate-tolerant crops. Uses of glyphosate-based plant protection products in third countries are evaluated by the competent authorities in those countries against the locally prevailing regulatory framework, but not against the criteria of Regulation (EC) No. 1107/2009. (...)." www.testbiotech.org/content/eu-commission-request-consider-impact-glyphosate-residues-feed-animal-health-february-2016

Consequently, EFSA should have requested the company to submit data from field trials with the highest dosage of the complementary herbicides that can be tolerated by the plants, including repeated spraying. The material derived from those plants should have been assessed in regard to organ toxicity, immune system responses and reproductive toxicity, also taking combinatorial effects with other plant components into account.

It is known that soybeans contain many biologically active substances, e.g. estrogens, allergens and anti-nutritional compounds, which may interact with trait-related characteristics and act as stressors. Changes in the composition of these components cannot only be triggered by the process of genetic engineering but also by interactions with the complementary herbicides. For example, Zobiole et al. (2012) and also Bøhn et al. (2014) found that glyphosate application can cause significant changes in soybean plant constituents. More specifically, Zobiole et al. (2012) applied glyphosate at three different dosages (800, 1200 and 2400 g/ha) which resulted in dose-correlated changes in plant agronomic performance and plant composition.

There are further relevant issues: for example, the potential impact on the intestinal microbiome also has to be considered. Such effects might be caused by the residues from spraying since glyphosate has been shown to have negative effects on the composition of the intestinal flora of cattle (Reuter et al., 2007), poultry (Shehata et al., 2013) and rodents (Mao et al., 2018).

In general, antibiotic effects and other adverse health effects might occur from exposure to a diet containing these plants that were not assessed under pesticide regulation. These adverse effects on health might be triggered by the residues from spraying with the complementary herbicide (see also van Bruggen et al., 2017). Furthermore, attention should be paid to the specific toxicity of the metabolites in the active ingredients of the pesticide that might occur specifically in the stacked event.

Whatever the case, both the EU pesticide regulation and the GMO regulation require a high level of protection for health and the environment. Thus, in regard to herbicide-resistant plants, specific assessment of residues from spraying with complementary herbicides must be considered to be a prerequisite for granting authorisation.

In regard to Cry toxins and their complex mode of action, EFSA only mentions two industry studies (Hammond et al., 2013; Koch et al., 2015) claiming that Bt toxins would only be active in targeted insects. However, these issues need to be assessed more thoroughly: as shown in publications, Bt toxins raise several questions in regard to feed and food safety:

- (1) There are several partially diverging theories about the exact mode of action of the Bt toxins at the molecular level (see Then, 2010; Hilbeck & Otto, 2015). Thus, it cannot be assumed a priori that the toxins are inert in regard to human and animal health as argued in risk assessment for food and feed carried out by Monsanto.
- (2) There are further uncertainties regarding the specificity of Bt toxins (Venter and Bøhn, 2016). Changes in specificity may emerge from structural modifications performed to render higher efficacy (see Hilbeck and Schmidt, 2006).
- (3) In addition, there are findings in mammalian species showing that Bt toxicity is a relevant topic for detailed health risk assessment: some Cry toxins are known to bind to epithelial cells in the intestines of mice (Vázquez-Padrón et al., 1999).

- (4) As far as potential effects on health are concerned, several publications (Thomas and Ellar 1983; Shimada et al., 2003; Mesnage et al., 2013; Huffman et al., 2004; Bondzio et al., 2013) show that Cry proteins may indeed have an impact on the health of mammals. For example, de Souza Freire et al., (2014) confirm hematotoxicity of several Cry toxins. Some of these effects seem to occur where there are high concentrations and tend to become stronger over longer periods of time.
- (5) Further, the toxicity of Bt toxins can be enhanced through interaction with other compounds, such as plant enzymes (Zhang et al., 2000, Zhu et al., 2007; Pardo-López et al., 2009), other Bt toxins (Sharma et al., 2004; Tabashnik et al., 2013; Bøhn et al. 2016, Bøhn 2018), gut bacteria (Broderick et al., 2009), residues from spraying with herbicides (Bøhn et al. 2016, Bøhn 2018) and other co-stressors (Kramarz et al., 2007; Kramarz et al., 2009; Khalique and Ahmed, 2005; Singh et al., 2007; Zhu et al., 2005; Mason et al., 2011; Reardon et al., 2004).

In this context, it is relevant that Bt toxins can persist in the gut to a much higher degree than has been assumed by EFSA. Chowdhury et al., (2003) and Walsh et al. (2011) have found that when pigs were fed with Bt maize, Cry1A proteins could frequently and successfully still be found in the colon of pigs at the end of the digestion process. This means that Bt toxins are not degraded quickly in the gut and can persist in larger amounts until digestion is completed; and that there is enough time for interaction between various food compounds. Especially in soybeans, compounds such as trypsin inhibitors, can delay the degradation of Bt toxins (Pardo-López et al., 2009) and can therefore cause higher exposure and render higher toxicity compared to experiments with the proteins in isolation. It has to be emphasised that the data presented on thermal or enzymatic degradation of the isolated proteins do not allow the assessment of the true persistence of the Bt toxins in the food chain.

Further, as far as the exposure of the food chain with Bt toxins is concerned, EFSA should have requested data on the overall combined exposure to Bt toxins caused by the introduction of Bt plants in the EU. Currently, there are around 40 events that produce Bt toxins authorised for import. The exposure stemming from these imports, taking into account maize gluten, should have been added to that of the stacked soybean assess exposure in a much more realistic scenario.

EU legal provisions such as Regulation 1829/2003 (as well as Implementing Regulation 503/2013) state that "any risks which they present for human and animal health and, as the case may be, for the environment" have to be avoided. Therefore, potential adverse effects that result from combinatorial exposure of various potential stressors need specification, and their assessment needs to be prioritised. We conclude that the health risk assessment as currently performed by EFSA for the stacked soybean is unacceptable. We propose that these plants are tested following the whole mixture approach, considering them to be "insufficiently chemically defined to apply a component-based approach" (EFSA, 2019d).

Despite all these open questions regarding potential health impacts, we are not aware of a single sub-chronic or chronic feeding study performed with whole food and feed derived from the stacked soybean. This observation is supported by a literature review carried out by the company that did not yield any peer-reviewed publication.

In conclusion, the EFSA opinion on the application for authorisation of the stacked soybean (EFSA, 2019a) cannot be said to fulfill assessment requirements of potential synergistic or antagonistic effects resulting from the combination of the transformation events in regard to toxicology.

As a result, the toxicological assessment carried out by EFSA is not acceptable.

Allergenicity

Implementing Regulation 503/2013 requests:

"In cases when known functional aspects of the newly expressed protein or structural similarity to known strong adjuvants may indicate possible adjuvant activity, the applicant shall assess the possible role of these proteins as adjuvants. As for allergens, interactions with other constituents of the food matrix and/or processing may alter the structure and bioavailability of an adjuvant and thus modify its biological activity."

"In accordance with the requirements of Articles 4 and 16 of Regulation (EC) No 1829/2003, the applicant shall ensure that the final risk characterisation clearly demonstrates that:

(a) the genetically modified food and feed has no adverse effects on human and animal health:"

However, EFSA did not request the applicant to provide data to verify whether the source of the transgene is allergenic. According to Santos-Vigil et al. (2018), the Bt toxin Cry1Ac can act as an allergen if ingested. The Bt toxin Cry1Ac was also used as a source for the synthesis of Cry1A.105 expressed in the stacked soybean. Therefore, the synthetically derived Cry1A.105 toxin produced in the soybean has structural similarity with Cry1Ac. If Cry1Ac is suspected of being an allergen, the source of Cry1A.105 has to be verified as allergenic and therefore investigated in detail.

The EU Commission initially noted that the Santos-Vigil et al. (2018) publication was relevant for the risk assessment of genetically engineered plants producing Bt toxins, and therefore requested the European Food Safety Authority (EFSA) for an assessment. However, EFSA (EFSA, 2018) came to the conclusion that the Santos-Vigil et al. (2018) publication does not provide any new information and suffers from methodological flaws. However, this EFSA opinion is based on a rather biased interpretation of existing publications, and it does not provide any evidence that the Santos-Vigil et al. (2018) findings are invalid or irrelevant (Moreno-Fierros et al., 2018).

In conclusion, the EFSA assessment of the stacked soybean cannot be said to fulfil the requirements for assessing allergenicity of the source of the transgene. The Santos-Vigil et al. (2018) publication has to be considered valid and not properly assessed by EFSA (Moreno-Fierros et al., 2018). In awareness of these findings, EFSA should have started with the hypothesis that the consumption of products derived from the soybean can trigger allergic reactions – and should therefore have requested empirical investigations.

Furthermore, there are several studies indicating that immune responses such as adjuvanticity in mammals are triggered by Bt toxins and have to be considered in this context. Studies with the Cry1Ac toxin (Moreno-Fierros et al., 2000; Vázquez-Padrón et al., 1999; Legorreta-Herrera et al., 2010; Jarillo-Luna et al. 2008; González-González et al., 2015; Ibarra-Moreno et al., 2014; Moreno-Fierros, 2007; Guerrero et al., 2004; Moreno-Fierros et al. 2013; Rubio-Infante et al. 2018) are especially relevant (for review also see Rubio-Infante et al. 2016).

In this context, it is relevant that Bt toxins can persist in the gut to a much higher degree than has been assumed by EFSA. Chowdhury et al., (2003) and Walsh et al. (2011) have found that when pigs were fed with Bt maize, Cry1A proteins could frequently and successfully still be found in the colon of pigs at the end of the digestion process. This means that Bt toxins are not degraded quickly in the gut and can persist in larger amounts until digestion is completed; and that there is enough time for interaction between various food compounds. Especially in soybeans, compounds such as trypsin inhibitors, can delay the degradation of Bt toxins (Pardo-López et al., 2009) and can

therefore cause higher exposure and render higher toxicity compared to experiments with the proteins in isolation. It has to be emphasised that the data presented on thermal or enzymatic degradation of the isolated proteins do not allow the assessment of the true persistence of the Bt toxins in the food chain.

Further, as far as the exposure of the food chain with Bt toxins is concerned, EFSA should have requested data on the overall combined exposure to Bt toxins caused by the introduction of Bt plants in the EU. Currently, there are already 30 events that produce Bt toxins authorised for import. The exposure stemming from these imports, taking into account maize gluten, should have been added to that of the stacked soybean assess exposure in a much more realistic scenario.

Given the fact that potential effects of Bt toxins on the immune system have meanwhile been discussed for many years (for overview see, for example, Then & Bauer-Panskus, 2017), and already around 40 GE crops events producing Bt toxins have been approved for the EU market, any further delay in resolving these crucial questions cannot be accepted. In accordance with EU Regulation 1829/2003, safety of whole food and feed has to be demonstrated before approval for import can be issued. Since this is not the case with the stacked soybean, the risk assessment is not conclusive and no market authorisation can be granted.

In summary, the EFSA assessment of the stacked soybean cannot be said to fulfill the requirements for assessing risks to the immune system.

Others

(1) From studying the statements of the experts from Member States (EFSA, 2019b), we have the impression that EFSA (2019a) is not aware of more recent publications showing a higher degree of horizontal gene transfer (HGT) than previously thought. Further, in their interpretation of the data, EFSA seems to be adopting a biased approach based on the assumption that no HGT should be expected.

In addition, given the fact that stacked events always show a higher overall amount of additionally inserted DNA, the statistical expectation of HGT involving this specific DNA needs more consideration. We conclude that the EFSA conclusions in regard to HGT to the intestinal gut of livestock and humans as well as the fate of the DNA in the environment will need further assessment.

(2) For monitoring and methods to identify the specific event, Implementing Regulation 503/2013 requests:

The method(s) shall be specific to the transformation event (hereafter referred to as 'event-specific') and thus shall only be functional with the genetically modified organism or genetically modified based product considered and shall not be functional if applied to other transformation events already authorised; otherwise the method cannot be applied for unequivocal detection/identification/quantification. This shall be demonstrated with a selection of non-target transgenic authorised transformation events and conventional counterparts. This testing shall include closely related transformation events.

However, no such method for identification was made available. Based on the information that is available, it will not be possible to distinguish the stacked event from a mixture of single parental events or stacked events that overlap with the actual stack.

If approval for import is given, the applicant has to ensure that post-market monitoring (PMM) is

developed to collect reliable information on the detection of indications showing whether any (adverse) effects on health may be related to GM food or feed consumption. Thus, the monitoring report should at very least contain detailed information on: i) actual volumes of the GE products imported into the EU; ii) the ports and silos where shipments of the GE products were unloaded; iii) the processing plants where the GE products was transferred to; iv) the amount of the GE products used on farms for feed; v) transport routes of the GE products. Environmental monitoring should be run in regions where viable material of the GE products such as kernels are transported, stored, packaged, processed or used for food/feed. In case of losses and spread of viable material (such as kernels), all receiving environments need to be monitored. Furthermore, environmental exposure through organic waste material, by-products, sewage or faeces containing GE products during or after the production process; and during or after human or animal consumption, should be part of the monitoring procedure (see also comments from experts of Member States, EFSA, 2019b).

(3) We agree with comments made by experts from Member States (EFSA 2019b), that the applicant should be asked to provide a detailed analysis of the fate of the Bt proteins in the environment and a quantitative estimate of subsequent exposure of non-target organisms.

Besides methods of detection, other methods for quantifying exposure to the insecticidal proteins need to be made publicly available in order to facilitate monitoring. Food and feed producers, farmers as well as experts dealing with environmental exposure (for example which waste material, spillage and manure) have to be able to gather independent information on their exposure to the toxins via independent laboratories. As yet, these methods are regarded as confidential business information and are not made available upon request by EFSA. Thus, the Commission should ensure that the relevant data are both publicly available and also reliable.

As existing evidence shows (Székács et al., 2011; Shu et al., 2018), the methods need to be carefully evaluated to ensure that the results are reliable, comparable and reproducible. Therefore, fully evaluated methods have to be published that allow the Bt concentration in the plants to be measured by independent scientists, as is the case for other plant protection compounds used in food and feed production. This is necessary to make sure that the environment as well as human and animals coming into contact with the material (for example, via dust, consumption or manure) are not exposed to higher quantities of Bt toxins than described in the application. But instead of requesting reliable testing methods, EFSA even refers to the insufficiency of the methods to explain unexpected and diverging results: "It is expected that variation in the protein quantification can occur in assays due to technical reasons". (EFSA, 2019b) This statement and the approach of EFSA cannot be accepted, because it is not in line with the high scientific standards as requested in GMO Regulation 1829/2003.

(4) Finally, in regard to the literature research, we do not agree with the way it was carried out. The review should take into account all publications on the parental plants and provide all relevant information regarding gene expression, findings from field trials and feeding studies. Further, monitoring data should be provided on imports of parental plants into the EU.

Environmental risk assessment

The EFSA (2019a) statement in regard to potential persistence of seeds and plants after spillage is not adequate:

"It is unlikely that the intended traits of soybean MON87751 x MON87701 x MON87708 x MON89788 will provide a selective advantage to soybean plants, except when they are exposed to dicamba- and/or glyphosate-containing herbicides or infested by insect pests that are susceptible to the Cry1A.105, Cry2Ab2 and/or Cry1Ac proteins. However, this fitness advantage will not allow the GM plant to overcome other biological and abiotic factors (described above). Therefore, the presence of the intended traits will not affect the persistence and invasiveness of the GM plant."

EFSA should reconsider this statement in the light of the findings of Fang et al (2018) which shows that there are unintended effects emerging from the production of the additional EPSPS enzymes in the plants. These findings make it necessary to request experimental data from the applicant regarding the real environmental persistence of the GE soybeans after spillage.

Conclusions and recommendations

The EFSA risk assessment cannot be accepted.

References

Avila-Vazquez, M., Difilippo, F.S., Lean, B.M., Maturano, E., Etchegoyen, A. (2018) Environmental exposure to glyphosate and reproductive health impacts in agricultural population of Argentina. J Environ Prot 9: 241–253. https://doi.org/10.4236/jep.2018.93016

Benbrook, C.M. (2016) Trends in glyphosate herbicide use in the United States and globally. Environ Sci Eur, 28: 3. https://doi.org/10.1186/s12302-016-0070-0

Bøhn, T., (2018) Criticism of EFSA's scientific opinion on combinatorial effects of 'stacked' GM plants. Food and Chemical Toxicology, 111: 268-274. https://doi.org/10.1016/j.fct.2017.11.023

Bøhn, T., Cuhra, M., Traavik, T., Sanden, M., Fagan, J., Primicerio, R. (2014) Compositional differences in soybeans on the market: glyphosate accumulates in Roundup Ready GM soybeans. Food Chem, 153: 207-215. https://doi.org/10.1016/j.foodchem.2013.12.054

Bøhn, T., Rover, C.M., Semenchuk, P.R. (2016) *Daphnia magna* negatively affected by chronic exposure to purified Cry-toxins. Food Chem. Toxicol., 91: 130-140. https://doi.org/10.1016/j.fct.2016.03.009

Bombardi, L.M. (2016) Pequeno ensaio cartográfico sobre o uso de agrotóxicos no Brasil. Laboratório de Geografia Agrária-USP, São Paulo. OpenEdition. https://journals.openedition.org/confins/12594

Bondzio, A., Lodemann, U., Weise, C., Einspanier, R. (2013) Cry1Ab Treatment has no effects on viability of cultured porcine intestinal cells, but triggers Hsp70 expression. PLOS ONE, 8(7): e67079. doi: 10.1371/journal.pone.0067079

Broderick, N.A., Robinson, C.J., McMahon, M.D., Holt, J., Handelsman, J., Raffa, K.F. (2009) Contributions of gut bacteria to Bacillus thuringiensis-induced mortality vary across a range of Lepidoptera. BMC Biol., 7: 11. https://doi.org/10.1186/1741-7007-7-11

Chowdhury, E.H., Kuribara, H., Hino, A., Sultana, P., Mikami, O., Shimada, N., Guruge, K.S., Saito, M., Nakajima, Y. (2003) Detection of corn intrinsic and recombinant DNA fragments and Cry1Ab protein in the gastrointestinal contents of pigs fed genetically modified corn Bt11. J. Anim. Sci., 81(10): 2546-2551. https://doi.org/10.2527/2003.81102546x

de Souza Freire, I., Miranda-Vilela, A.L., Barbosa, L.C.P., et al. (2014) Evaluation of cytotoxicity, genotoxicity and hematotoxicity of the recombinant spore-crystal complexes Cry1Ia, Cry10Aa and Cry1Ba6 from Bacillus thuringiensis in swiss mice. Toxins, 6: 2872-2885. https://doi.org/10.3390/toxins6102872

EFSA (2018) Relevance of new scientific information (Santos-Vigil et al., 2018) in relation to the risk assessment of genetically modified crops with Cry1Ac. EFSA supporting publication: EN-1504. 13 pp. doi:10.2903/sp.efsa.2018.EN-1504. https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1504

EFSA (2019a) Assessment of genetically modified soybean MON 87751 × MON 87701 × MON 87708 × MON 89788 for food and feed uses, under Regulation (EC) No 1829/2003 (application EFSA-GMO-NL-2016-128). EFSA Journal, 17(11): e05847. https://doi.org/10.2903/j.efsa.2019.5847

EFSA (2019b) Application EFSA-GMO-NL-2016-128, Comments and opinions submitted by Member States during the three-month consultation period, Register of Questions, http://registerofquestions.efsa.europa.eu/roqFrontend/questionsListLoader?unit=GMO

EFSA (2019c) Review of the existing maximum residue levels for glyphosate according to Article 12 of Regulation (EC) No 396/2005 – revised version to take into account omitted data. EFSA Journal 2019; 17(10): 5862, 211 pp. https://doi.org/10.2903/j.efsa.2019.5862

EFSA (2019d) Guidance on harmonised methodologies for human health, animal health and ecological risk assessment of combined exposure to multiple chemicals. EFSA Journal, 17(3): 5634, 77 pp. https://doi.org/10.2903/j.efsa.2019.5634

Fang, J., Nan, P., Gu, Z. et al. (2018) Overexpressing exogenous 5-enolpyruvylshikimate-3-phosphate synthase (EPSPS) genes increases fecundity and auxin content of transgenic arabidopsis plants. Front Plant Sci, 9: 233. https://www.frontiersin.org/articles/10.3389/fpls.2018.00233

González-González, E., García-Hernández A.L., Flores-Mejía, R., López-Santiago, R., Moreno-Fierros L. (2015) The protoxin Cry1Ac of *Bacillus thuringiensis* improves the protection conferred by intranasal immunization with *Brucella abortus* RB51 in a mouse model. Vet. Microbiol., 175: 382-388. http://dx.doi.org/10.1016/j.vetmic.2014.11.021

Guerrero, G.G. & Moreno-Fierros, L. (2007) Carrier potential properties of *Bacillus thuringiensis* Cry1A toxins for a diphtheria toxin epitope, Scandinavian Journal of Immunology, 66(6): 610-618. http://dx.doi.org/10.1111/j.1365-3083.2007.01992.x

Guerrero, G.G., Dean, D.H., Moreno-Fierros, L. (2004) Structural implication of the induced immune response by *Bacillus thuringiensis* cry proteins: role of the N-terminal region, Molecular Immunology, 41(12): 1177-1183. http://dx.doi.org/10.1016/j.molimm.2004.06.026

Hammond, B., Kough, J., Herouet-Guicheney, C., Jez, J.M. (2013) Toxicological evaluation of proteins introduced into food crops. Critical Reviews in Toxicology, 43 (Suppl.2): 25-42. https://www.tandfonline.com/doi/full/10.3109/10408444.2013.842956

Hilbeck, A., & Schmidt, J.E. (2006) Another view on Bt-proteins-how specific are they and what else might they do. Biopesticides International, 2(1): 1-50. https://www.researchgate.net/publication/238550409 Another View on Bt Proteins - How Specific are They and What Else Might They Do

Hilbeck, A., Otto, M. (2015) Specificity and Combinatorial effects of *Bacillus Thuringiensis* Cry toxins in the context of GMO environmental risk assessment. Front Environ Sci, 3: 71. doi: 10.3389/fenvs.2015.00071

Huffman, D.L., Abrami, L., Sasik, R., et al. (2004) Mitogen-activated protein kinase pathways defend against bacterial pore-forming toxins. Proc Natl Acad Sci U S A, 101(30): 10995-11000. doi: 10.1073/pnas.0404073101

Ibarra-Moreno, S., García-Hernández, A.L., Moreno-Fierros L. (2014) Coadministration of protoxin Cry1Ac from *Bacillus thuringiensis* with metacestode extract confers protective immunity to murine cysticercosis. Parasite Immunol. 36(6): 266–270. http://dx.doi.org/10.1111/pim.12103

Jarillo-Luna, A., Moreno-Fierros L., Campos-Rodríguez R., Rodríguez-Monroy, M.A., Lara-Padilla, E., Rojas-Hernández, S. (2008) Intranasal immunization with *Naegleria fowleri* lysates and Cry1Ac induces metaplasia in the olfactory epithelium and increases IgA secretion. Parasite Immunol., 30: 31-38. http://dx.doi.org/10.1111/j.1365-3024.2007.00999.x

Khalique, F., Ahmed, K. (2005) Compatibility of bio-insecticide with chemical insecticide for management of *Helicoverpa armigera* Huebner. Pak. J. Biol. Sci. 8 (3), 475–478. https://doi.org/10.3923/pjbs.2005.475.478

Kleter, G.A., Unsworth, J.B., Harris, C.A. (2011) The impact of altered herbicide residues in transgenic herbicide-resistant crops on standard setting for herbicide residues. Pest Manag Sci, 67(10): 1193-1210. doi: 10.1002/ps.2128

Koch, M.S., Ward, J.M., Levine, S.L., et al. (2015) The food and environmental safety of Bt crops. Frontiers in Plant Science, 6: 283. https://www.frontiersin.org/articles/10.3389/fpls.2015.00283

Kramarz, P., de Vaufleury, A., Gimbert, F., Cortet, J., Tabone, E., Andersen, M.N., Krogh, P.H. (2009) Effects of Bt-maize material on the life cycle of the land snail *Cantareus aspersus*. Appl. Soil Ecol., 42(3): 236-242. https://doi.org/10.1016/j.apsoil.2009.04.007

Kramarz, P.E., De Vaufleury, A., Carey, M. (2007) Studying the effect of exposure of the snail *Helix aspersa* to the purified Bt toxin, Cry1Ab. Applied Soil Ecology, 37: 169-172. https://doi.org/10.1016/j.apsoil.2007.06.006

Legorreta-Herrera, M., Oviedo Meza, R., Moreno-Fierros L. (2010) Pretreatment with Cry1Ac protoxin modulates the immune response, and increases the survival of *plasmodium* -infected CBA/Ca mice, J Biomed Biotechnol,2010: 198921. http://dx.doi.org/10.1155/2010/198921

Mao, Q., Manservisi, F., Panzacchi, S., Mandrioli, D., Menghetti, I., Vornoli, A., Bua, L., Falcioni, L., Lesseur, C., Chen, J., Belpoggi, F., Hu, J. (2018) The Ramazzini Institute 13-week pilot study on glyphosate and Roundup administered at human-equivalent dose to Sprague Dawley rats: effects on the microbiome. Environmental Health, 17: 50. https://doi.org/10.1186/s12940-018-0394-x

Mason, K.L., Stepien, T.A., Blum, J.E., Holt, J.F., Labbe, N.H., Rush, J.S., Raffa, K.F., Handelsman, J. (2011) From commensal to pathogen: translocation of *Enterococcus faecalis* from the midgut to the hemocoel of *Manduca sexta*. mBio 2(3): e00065-00011. https://doi.org/10.1128/mBio.00065-11

Mesnage, R., Clair, E., Gress, S., et al. (2013) Cytotoxicity on human cells of Cry1Ab and Cry1Ac Bt insecticidal toxins alone or with a glyphosate-based herbicide. Journal of Applied Toxicology 33(7): 695-699. doi: 10.1002/jat.2712

Monsanto (2017) Roundup PowerMaxII product label. Greenbook. https://assets.greenbook.net/16-50-19-15-03-2018-63045R2-9_PMAXII_SpecimenLabel_2017.pdf

Moreno-Fierros, L., García N., Gutiérrez, R., López-Revilla, R., Vázquez-Padrón, R.I. (2000) Intranasal, rectal and intraperitoneal immunization with protoxin Cry1Ac from *Bacillus thuringiensis* induces compartmentalized serum, intestinal, vaginal and pulmonary immune responses in Balb/c mice. Microbes Infect., 2(8): 885–890. http://dx.doi.org/10.1016/S1286-4579(00)00398-1

Moreno-Fierros, L., García-Hernández, A.L., Ilhuicatzi-Alvarado, D., Rivera-Santiago, L., Torres-Martínez, M., Rubio-Infante N., Legorreta-Herrera, M. (2013) Cry1Ac protoxin from *Bacillus thuringiensis* promotes macrophage activation by upregulating CD80 and CD86 and by inducing IL-6, MCP-1 and TNF-α cytokines, Int. Immunopharmacol., 17(4): 1051-1066, http://dx.doi.org/10.1016/j.intimp.2013.10.005

Moreno-Fierros, L., Santos-Vigil, K., Ilhicatzi-Alvarado, D. (2018) Response to assessment of the elevance of new scientific information (Santos-Vigil et al., 2018) in relation to the risk assessment of genetically modified crops with Cry1Ac of European Food Safety Authority (EFSA). www.testbiotech.org/node/2304

Pardo-López, L., Muñoz-Garay, C., Porta, H., Rodríguez-Almazán, C., Soberón, M., Bravo, A. (2009) Strategies to improve the insecticidal activity of Cry toxins from *Bacillus thuringiensis*. Peptides, 30(3): 589-595. https://www.sciencedirect.com/science/article/pii/S0196978108003264

Reardon, B.J., Hellmich, R.L., Sumerford, D.V., Lewis, L.C. (2004) Growth, development, and survival of *Nosema pyrausta*-infected european corn borers (*Lepidoptera*: Crambidae) reared on meridic diet and Cry1Ab. J. Econ. Entomol., 97(4): 1198–1201. https://doi.org/10.1093/jee/97.4.1198

Reuter, T., Alexander, T.W., Martínez, T.F., McAllister, T.A. (2007) The effect of glyphosate on digestion and horizontal gene transfer during in vitro ruminal fermentation of genetically modified canola. J Sci Food Agric, 87(15): 2837-2843. doi: 10.1002/jsfa.3038

Reuter, W. (2015) Toxicology of glyphosate, isoxaflutole, dicamba and possible combination effects. Testbiotech. https://www.testbiotech.org/content/toxicology-glyphosate-isoxaflutole-dicamba-and-possible-combination-effects

Rubio Infante, N., & Moreno-Fierros, L. (2016) An overview of the safety and biological effects of Bacillus thuringiensis Cry toxins in mammals. Journal of Applied Toxicology, 36(5): 630-648. http://onlinelibrary.wiley.com/doi/10.1002/jat.3252/full

Rubio-Infante, N., Ilhuicatzi-Alvarado, D., Torres-Martínez, M., et al. (2018) The macrophage activation induced by *Bacillus thuringiensis* Cry1Ac protoxin involves ERK1/2 and p38 pathways and the interaction with cell-surface-HSP70. Journal of Cellular Biochemistry 119(1): 580–598. doi: 10.1002/jcb.26216

Santos-Vigil, K.I., Ilhuicatzi-Alvarado, D., García-Hernández, A.L., Herrera-García, J.S., Moreno-Fierros, L. (2018) Study of the allergenic potential of *Bacillus thuringiensis* Cry1Ac toxin following intra-gastric administration in a murine model of food-allergy. International immunopharmacology, 61: 185-196. https://www.sciencedirect.com/science/article/pii/S1567576918302467

Sharma, H.C., Sharma, K.K., Crouch, J.H., (2004) Genetic transformation of crops for insect resistance: potential and limitations. Crit. Rev. Plant Sci. 23(1), 47–72. https://doi.org/10.1080/07352680490273400

Shehata, A.A., Schrödl, W., Aldin, A.A., et al. (2013) The effect of glyphosate on potential pathogens and beneficial members of poultry microbiota in vitro. Curr Microbiol 66(4): 350–358. doi: 10.1007/s00284-012-0277-2

Shimada, N., Kim, Y.S., Miyamoto, K., Yoshioka, M., Murata, H. (2003) Effects of *Bacillus thuringiensis* Cry1Ab toxin on mammalian cells. J Vet Med Sci, 65(2): 187-191. https://www.jstage.jst.go.jp/article/jvms/65/2/65_2_187/_article/-char/ja/

Shu, Y., Romeis, J., Meissle, M. (2018) No interactions of stacked Bt maize with the non-target aphid Rhopalosiphum padi and the spider mite Tetranychus urticae. Front. Plant Sci., 9: 39. https://doi.org/10.3389/fpls.2018.00039

Singh, G., Rup, P.J., Koul, O. (2007) Acute, sublethal and combination effects of azadirachtin and *Bacillus thuringiensis* toxins on *Helicoverpa armigera* (Lepidoptera: Noctuidae) larvae. Bull. Entomol. Res., 97(4): 351-357. https://doi.org/10.1017/S0007485307005019

Székács, A., Weiss, G., Quist, D., Takács, E., Darvas, B., Meier, M., Swain, T., Hilbeck, A. (2012) Inter-laboratory comparison of Cry1Ab toxin quantification in MON 810 maize by enzyme-immunoassay. Food Agric. Immunol. 23(2): 99-121. https://doi.org/10.1080/09540105.2011.604773

Tabashnik, B.E., Fabrick, J.A., Unnithan, G.C., Yelich, A.J., Masson, L., Zhang, J., Bravo, A., Soberón, M. (2013) Efficacy of genetically modified Bt toxins alone and in combinations against pink bollworm resistant to Cry1Ac and Cry2Ab. PloS one 8(11): e80496. https://doi.org/10.1371/journal.pone.0080496

Then, C. (2010) Risk assessment of toxins derived from Bacillus thuringiensis – synergism, efficacy, and selectivity. Environ Sci Pollut Res 17(3): 791–797. https://link.springer.com/article/10.1007/s11356-009-0208-3 Then, C., & Bauer-Panskus, A. (2017) Possible health impacts of Bt toxins and residues from spraying with complementary herbicides in genetically engineered soybeans and risk assessment as performed by the European Food Safety Authority EFSA. Environmental Sciences Europe, 29(1): 1. https://enveurope.springeropen.com/articles/10.1186/s12302-016-0099-0

Thomas, W.E. & Ellar, D.J. (1983) *Bacillus thuringiensis* var israelensis crystal delta-endotoxin: effects on insect and mammalian cells in vitro and in vivo. Journal of Cell Science, 60(1): 181-197. http://jcs.biologists.org/content/60/1/181.short

Trtikova, M., Wikmark, O.G., Zemp, N., Widmer, A., Hilbeck, A. (2015) Transgene expression and Bt protein content in transgenic Bt maize (MON810) under optimal and stressful environmental conditions. PloS one, 10(4): e0123011.

http://journals.plos.org/plosone/article?id=10.1371/journal.pone.0123011

United States Department of Agriculture (USDA), National Agricultural Statistics Service (NASS) (2019) Agricultural chemical use program.

http://www.nass.usda.gov/Surveys/Guide to NASS Surveys/Chemical Use/index

van Bruggen, A.H.C., He, M.M., Shin, K., Mai, V., Jeong, K. C., Finckh, M.R., Morris, J.G. (2017) Environmental and health effects of the herbicide glyphosate. Science of The Total Environment, 616: 255-268. https://www.sciencedirect.com/science/article/pii/S0048969717330279

Vázquez-Padrón, R.I., Moreno-Fierros, L., Neri-Bazán, L., de la Riva, G.A., López-Revilla, R. (1999) Intragastric and intraperitoneal administration of Cry1Ac protoxin from *Bacillus thuringiensis* induces systemic and mucosal antibody responses in mice. Life Sciences, 64(21): 1897–1912. doi: 10.1016/S0024-3205(99)00136-8

Venter, H.J., Bøhn, T. (2016) Interactions between Bt crops and aquatic ecosystems: A review. Environ Toxicol Chem, 35: 2891-2902. https://doi.org/10.1002/etc.3583

Walsh, M.C., Buzoianu, S.G., Gardiner, G.E., Rea, M.C., Gelencsér, E., Jánosi, A., Epstein, M.M., Ross, R.P., Lawlor, P.G. (2011) Fate of transgenic DNA from orally administered Bt MON810 maize and effects on immune response and growth in pigs. PloS one, 6(11): e27177. https://doi.org/10.1371/journal.pone.0027177

Zhang, J., Wang, C., Qin, J. (2000) The Interactions between soybean trypsin inhibitor and δ-endotoxin of *Bacillus thuringiensis* in *Helicoverpa armigera* larva. Journal of Invertebrate Pathology, 75: 259-266. https://doi.org/10.1006/jipa.2000.4936

Zhu, Y.C., Abel, C.A., Chen, M.S. (2007) Interaction of Cry1Ac toxin (*Bacillus thuringiensis*) and proteinase inhibitors on the growth, development, and midgut proteinase activities of the bollworm, *Helicoverpa zea*. Pesticide Biochemistry and Physiology, 87: 39-46. https://doi.org/10.1016/j.pestbp.2006.05.004

Zhu, Y.C., Adamczyk, J.J., West, S. (2005) Avidin, a potential biopesticide and synergist to *Bacillus thuringiensis* toxins against field crop insects. J. Econ. Entomol. 98(5): 1566-1571. https://doi.org/10.1093/jee/98.5.1566

Zobiole, L.H.S., Kremer, R.J., de Oliveira, Jr. R.S., Constantin, J. (2012) Glyphosate effects on photosynthesis, nutrient accumulation, and nodulation in glyphosate resistant soybean. J Plant Nutr Soil Sci, 175(2): 319–330. https://doi.org/10.1002/jpln.201000434